

Early Life Exposure to Outdoor Air Pollution: Effect on Child Health in India

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Abstract

This paper examines effect of outdoor air pollution on child health in India by combining satellite PM2.5 data with geo-coded Demographic and Health Survey of India(2016). We use an instrumental variable strategy for identification as local pollution levels may be endogenous due to local household behavioural choices like participation in local fuel wood market, burning crop residue etc which are not observed in survey data. Our identification strategy relies on use of upwind biomass burning events in neighbouring areas to identify the effect of air pollution on child health. Our results indicate that one standard deviation increase in exposure to pollution during first trimester lowers Height-for-age (by 6.7 percent) and Weight-for-age (by 7.8 percent); the effect is prominent for poor people and Northern states of India which have higher incidence of such events.

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Effect of Early Life Exposure to Outdoor Air Pollution on Child Health in India ¹

[PRELIMINARY DRAFT]

1 Introduction

Pollution in any form, whether it be air or water, poses an environmental risk to the health of the exposed population. Literature from both developing and developed nations indicates the adverse health effects that air (Soo and Pattnayak, 2019 & Chay and Greenstone, 2003) and water pollution (Brainerd and Menon, 2014) have on children. In spite of enormous evidence about ill effects of pollution, there is little effective management or regulation to curtail activities which contribute to high levels of pollution in developing nations.

According to WHO global air pollution database, out of the 15 most polluted cities in the world, 14 belong to India. Another recently published report by Health Effects Institute on air pollution in India (2018) reports that air pollution was responsible for 1.1 million deaths in India in 2015. The major contributors to air pollution in India are household burning emissions, coal combustion, agricultural burning and transport. In the absence of effective pollution regulatory policies, air pollution levels have reached alarming levels in various parts of India (Greenpeace, 2017). This warrants a closer look at the air pollution problem from the standpoint of welfare of the younger generation currently being exposed to harmful pollutants with possible long-lasting effect on their health.

This paper examines the effect of outdoor air pollution on child health in India. Particularly we study the effect of early life exposure to air pollution (as measured by PM 2.5) on children's weight and height measures. We use gridded satellite data on PM2.5 and

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GPS locations of sampled clusters in Demographic Health Survey (DHS, 2015-16 round for India) to produce rich geo-spatial information about local pollution levels in the place of conception(residence) of a child. In an empirical exercise which causally links child health to local pollution levels, household income and behavioural choices are omitted variables which make local pollution levels endogenous. We use neighbouring (or non-local) fire-events as an instrument for local pollution levels as they are not related to household behavioural choices or local economic activity, but they affect local pollution levels as smoke and pollutants from these neighbouring fire events can travel long distances.

Our analysis shows that air pollution negatively affects children's health. Exposure to air pollution during the first trimester decreases both Height-for-age (stunting measure) and Weight-for-age (underweight measure) for children aged below five years. The effect is prominent for poorer households, with Northern states being more vulnerable due to high pollution levels in the area.

Evaluating this link between poor air quality and child health is important as many regions in India have very high pollution levels which breach the safe standards often. If stunting is affected by early life exposure to pollution then it can have long-lasting effect on life earnings of a child due to poor cognition and also increase the vulnerability towards hypertension and diabetes. Thus, failure to curb air pollution problem in India adversely affects the human capital of the country in the long run.

The literature linking air pollution to child health has mostly focused on child mortality. We add analysis towards child's growth indicators conditional on child's survival. Most of the studies for developing nations lack rich geo-spatial data for analysis and rely on pollution measures which are estimated for a large area (and hence are riddled with measurement error). To the best of our knowledge, this is the first study for India which addresses the endogeneity issues present while studying the link between child health with local pollution levels.

The paper follows the following structure. The next section provides a literature overview of the effect of pollution on child health. Section 3 describes the various datasets that we use in our analysis. The next section presents the empirical methodology that we follow and

it is followed by results in Section 5. Lastly, Section 6 concludes with an estimate of the extent of the problem and discusses current state of policies regarding air pollution in India.

2 Previous Literature

Our work is motivated by the “fetal origins” hypothesis (Douglas and Currie, 2011), which states that the *in-utero* period of a child critically determines mortality outcomes, disease prevalence and future health outcomes, abilities and earnings. Fetal growth, if restricted, can negatively affect future outcomes. The biological link between fetal growth and air pollution has not been documented in the literature, but it is mediated by placental growth which determines supply of oxygen and nutrients to the fetus. Exposure to pollution would affect placental function which can be impacted by inflammation caused by maternal infection. Additionally pollution is known to cause epigenetic changes (interaction between our genes and environment which can cause DNA methylation, which regulates gene expression) which could affect fetal growth as well. A recent paper by Chakrabarti et. al (2019) has shown how exposure to biomass burning (which causes pollution) affects respiratory health in adults as well as children. This link shows how mothers can possibly be affected during the pregnancy time due to exposure to pollution.

The intrauterine period has been the focus of many studies in economics literature which has linked occurrence of early life shocks to multiple outcomes. Early life shocks studied in economics literature include incidence of a) disastrous events (like famines, war, drought); b) nutritional shocks (like introduction of iodised salt, pregnancy during Ramadan) and c) pollution (air or water). Currie and Vogl (2013) provide a review of these early life shocks (a and b) on various outcomes; broadly summarised these shocks negatively affect adult cognition, years of schooling, literacy status, adult height and stunting measures; and increase the likelihood of presence of birth defects, prevalence of heart disease and obesity.

The focus of our study is in-utero exposure to air pollution and Currie et. al (2014) reviews landmark studies which have been done in this area. Most of these studies are from developed nations with few exceptions. Similar to previous studies, a major part of

the literature focuses on learning outcomes (test-scores) and earnings which are negatively affected due to in-utero exposure to pollution (Bharadwaj et al., 2013; Isen et al., 2013 & Sanders, 2012).

The strand of literature which is most relevant for our study has mainly looked at the effect of in-utero or early life exposure to air-pollution on infant mortality and birth weight. Few papers in this area have used natural experiments to causally identify the effect of air pollution on infant survival, for example, Chay and Greenstone (2003a and 2003b) use introduction of environmental regulations under Clean Air Act, 1970 and recession in 1981-82 in United States to show that reduction in pollution levels led to reduction in infant mortality. Currie and Walker (2011) show that introduction of congestion-reducing automated toll payment systems in United States (which reduced number of idle vehicles emitting harmful pollutants) reduced pre-mature and low birth-weight births. Currie and Neidell (2005) use spatial and temporal variation in CO levels to analyse the effect of CO levels on infant mortality.

The paper by Greenstone and Hanna (2014) is one of the landmark studies which analyses the effect of water and air pollution regulation policies on infant mortality in a developing nation context (India). Another study from a developing nation includes Foster et al. (2009) which uses Mexico's clean industry certification program to study its effect on pollution (we use a similar measure of pollution i.e. satellite data on Aerosol Optical Depth to infer PM_{2.5} levels) and resulting respiratory related infant deaths. Wildfires and their negative health effects (like increase in infant mortality, reported asthma cases, pre-term births etc) have also been studied in context of Indonesian wildfire of 1997 (Jayachandran, 2009; Rukumnuaykit, 2003; Kunii et al., 2002; Frankenberg et al., 2005 & Barber and James, 2000), California wildfires (Holstius et al., 2012) and Australian wildfires (O'Donnell and Behie, 2015). Two recent papers assess the effect of in-utero exposure to biomass burning events and pollution Vogl & Rangel (2016) and Soo & Pattanayak (2019) on birth weight and long-term health outcomes like adult height respectively. These two papers come closest to our paper as we also explore the link between in-utero exposure to pollution and child health but our sample is much bigger than the Indonesian study and we focus on solving the endogeneity problem

in our paper rather than focusing on reduced form effect of biomass burning events on child health.

Additionally two studies on India analyse the effect of water pollution on child health. Brainerd and Menon (2014) have focused on use of fertilisers in India during crop sowing season which increases concentration of harmful chemicals in water. They find that exposure to these pollutants during the month of conception increases infant mortality and reduces Height-for-age and Weight-for-age for children. Do et al. (2018) have shown that regulation targeting industrial pollution in the Ganga River led to reduction in water pollution levels and infant death. In case of India, literature has shown how air pollution can affect infant mortality but no study has focused on post-natal growth outcomes for surviving children. Also due to lack of local air pollution monitoring systems which cover the entire nation, there has been no study which calculates local pollution levels and assesses its effect on health outcomes.

3 Data

3.1 Demographic Data

The demographic data we use comes from the Demographic and Health Survey (Round-4 for 2015-16) for India. DHS-IV contains detailed information about birth history of each woman who was interviewed. The latest round of this survey sampled 601,509 households and interviewed 0.7 million eligible ² women in the age group 15-49. Further anthropometric measures of health were collected for 0.22 million children of age five years and below. The DHS sample is a stratified two-stage sample and the primary sampling units (PSUs or clusters) correspond to villages in rural areas and blocks in urban areas. The DHS-IV comprises of around 28526 clusters with GIS information on almost all clusters³. To hide the identity of the village (block in urban areas) all clusters were displaced by five kilometres (two kilometres for urban clusters), with one percent of the clusters being displaced by as

² Eligible women - married or unmarried women of reproductive age.

³ 131 clusters have no GIS information.

much as 10 kilometres. We account for this displacement when we discuss our identification strategy in the next section.

Our focus is on in-utero exposure to pollution for which we need the location and time of conception. To measure in-utero exposure to pollution we use the birth history of every child ever born to a woman. We use the location of the cluster, birth date and pregnancy duration of a child to impute exposure to pollution during the first trimester ⁴. We make an important assumption that the place of stay of mother when the child was in-utero is same as the same of current residence of a child ⁵.

We measure long-term impact by using anthropometric measures like Height-for-age and Weight-for-age (WHO standard z-scores) for children. While estimating the effect of in-utero exposure to pollution on measures of child health we also control for child characteristics like birth order, age and gender. As informed by previous literature, mother's characteristics can also affect child's health and hence we control for the age at which she had the child and her education level. DHS data also contains information about source of water, method of cooking and toilet facility of the dwelling, all of these characteristics have important bearing on health of a child and we include them in our analysis as well. Additional controls for gender and age of head of household are also used in our analysis.

We provide summary statistics of our analysis sample in Table 1. 52 percent of children in our estimation sample ⁶ are males with mean age around 29 months (2.5 years old) and the mean birth order of children is 2.2. The birth order is slightly lower for children in South India. The average age at which mothers have children is 24.5 years. Mothers had on an average 6.2 years of education (with mothers in South India having more years of education (7.8) than North Indian mothers (5.8)) and 97 percent of fathers are literate. Three-fourth of our sample consists of rural households and a similar proportion of households report their

⁴ We also construct separate measures of exposures to pollution for other trimesters and first three months after birth.

⁵ This assumption is a standard assumption which is employed by many papers which used DHS data for analysis. In our sample the mean number of years for which the interviewed family has stayed at the place of residence is around 15 years. See Brainerd & Menon, 2014 for reference.

⁶ Details about estimation sample discussed in pollution section later.

religion to be Hindu. The mean household size for our estimation sample is 6.5. 88 percent of the households are headed by a male member and the average age of household head is 44.5 years. 85 percent of the households have an electricity connection, but only 23 percent of the households use piped water as their source of drinking water, 28 percent of our sample uses clean source of cooking fuel like LPG or bio-gas and the mean open defecation rate in a cluster is 43 percent.

The mean Height-for-age (HFA) and Weight-for-age (WFA) z-score for our sample is -1.46 and -1.52 respectively (mean weight-for-height is -0.97 for our sample). Children from South India have much better HFA and WFA as compared to North Indian children. Height-for-age is a measure of stunting and it represents the effect of early life shocks that a child receives. Stunting generally occurs before age two and its effects are largely irreversible. It is associated with an underdeveloped brain, with long-lasting harmful consequences, including diminished mental ability and learning capacity, poor school performance in childhood, reduced earnings and increased risks of nutrition-related chronic diseases such as diabetes, hypertension, and obesity in future. Weight-for-age (underweight measure) reflects body mass relative to chronological age. It is influenced by both the height of the child (height-for-age) and his or her weight (weight-for-height). Deaton and Dreze (2009) advocate the use of Weight-for-age as the health status indicator for children as its a comprehensive measure which captures both stunting and wasting.

3.2 Pollution Data

In India, ground-based pollution measurement started post 2009 under the National Ambient Air Quality monitoring program maintained by Central Pollution Control Board. The network has slowly expanded to around 90 sites across 35 cities over the years, which leaves majority of India unmonitored ⁷. Amongst these cities, only Delhi has greater than 20 monitoring sites while most other cities have a single monitoring site. Furthermore, most of the sites do not have continuous temporal data. We use PM2.5 as our measure of pollution

⁷ India has around 600 ground based monitors to cover the entire country with only 148 monitors which capture PM2.5 for the entire country.

which is a correlate of other pollutants (like NO₂, SO₂, CO) which are not captured in our analysis. To address the paucity in ground-based pollution data in India, we estimate PM_{2.5} exposure using satellite data (van Donkelaar et al., 2010). We convert Aerosol Optical Data (AOD) retrieved at 0.5 x 0.5 degree resolution from Multiangle Imaging SpectroRadiometer (MISR) to PM_{2.5} data (Liu et al, 2004; Kahn and Gaitley 2015; Dey et al 2010) using a spatially and temporally heterogeneous conversion factor (Dey et al., 2012). The data was statistically downscaled at 0.1 x 0.1 degree resolution (using spline interpolation) PM_{2.5}. The PM_{2.5} data after processing is available at monthly frequency with a 0.1 * 0.1 degree resolution (10km*10km grid).

We explore the spatial variation in PM_{2.5} by plotting a heat map in Figure 1. We plot mean annual PM_{2.5} (average over monthly data for years 2010 to 2016) for each district of India. As the figure shows, the Northern region of the country is severely impacted by high and dangerous levels of pollution, especially the states which lie in Indo-Gangetic plains (Punjab, Haryana, Uttar Pradesh, Bihar) have the highest levels of pollution. On the other hand the Southern part of the country has much lower levels of pollution as shown by the lighter shades in heat map. The WHO guideline for maintaining safe standards of pollution recommends a threshold of mean annual pollution levels of 10ug/m³. Other standards include WHO-IT1 which is 35ug/m³, WHO-IT2 is 25ug/m³ and WHO-IT3 is 15ug/m³. The Indian National Ambient Air Quality Standards (NAAQS) is 40ug/m³. For our estimation sample we observe in Table 1 that the mean level of PM_{2.5} for is between 55ug/m³ & 60ug/m³ during all critical windows of development for children from North India while the corresponding level of pollution exposure is much lower for children from South India, that is around 35ug/m³ for all windows.

We use cluster location from DHS data and calculate mean PM_{2.5} in the 75km radius for each month since the time of conception. We use these monthly pollution measures to construct trimester level pollution exposure by calculating mean PM_{2.5} for three month periods. Our estimation sample is constrained by the availability of PM_{2.5} data as we only keep those children in our estimation sample for whom the pollution measure for each month in the first trimester is present. The missing PM_{2.5} are due to missing satellite retrievals

due to cloud covers. In the appendix (Table A1) we show that our outcome variables along with our control variables are very similar between the estimation sample and out-sample (with missing PM2.5 information).

We now link exposure to pollution during the first trimester with anthropometric measure (Height-for-age z-score) for children in Figure 2. The descriptive graph is a bin-scatter plot which shows a negative relationship between Height-for-age and exposure to pollution during first trimester, we explore this relationship empirically in greater detail in Section 4.

4 Empirical Model

We seek to investigate whether early life exposure to outdoor pollution during first trimester has an impact on future child health. We measure child health by Height-for-age (z-scores) and Weight-for age (z-scores). For ease of exposition we focus on Height-for-age and exposure in first trimester as early life exposure to pollution. The logic of our argument is same for all other outcomes and exposure windows. Formally, we estimate the following empirical model:

$$H_{icdmt} = \theta_1 PM_{cdmt} + \beta X_{icdmt} + \gamma_c + \delta_t + \lambda_m + \rho_{dt}^1 + \rho_{mt}^2 + \varepsilon_{icdmt} \quad (1)$$

Our main outcomes of interest (H_{icdmt}) are z-score for Height-for-age (stunting measure) and Weight-for-age (underweight measure) for child i who was conceived in month m , year t in cluster c belonging to district d . We also control for other confounding factors in the vector X_{icdmt} which includes gender, birth order and age of child, mother's and father's educational status, mother's age at birth, age and gender of household head, dummy for whether household has piped water, has clean cooking source, whether household practices open defecation and the fraction of households who practice open defecation in the cluster (excluding self). Since all children in our sample are aged five or below, we use the assumption that these controls have not changed a lot over time (i.e. from the time of conception to the time when they were surveyed).

The main variable of interest is PM_{cdmt} which captures the mean PM2.5 in the 75km

radius during first trimester for a child. Different clusters (villages or blocks) can have different levels of development (health infrastructure) which can affect health of a child hence we include cluster fixed effects in our specification. We also include month and year fixed effects to account for systematic effects related to season and year. We also remove any omitted variables that are related to a district in any particular year as well as any seasonality effect specific to a month of a particular year by including district into year, ρ_{dt}^1 and month into year, ρ_{mt}^2 fixed effects. The inclusion of these fixed effects means that the variation that remains is the spatial variation pollution across clusters and temporal variation within a year for a district.

While our estimation exercise removes systematic variation using various fixed effects but endogeneity concerns still remain. These endogeneity concerns arise as the *local* residential area for a household corresponds to the region of economic activity that a household depends on and also affects based on its behavioural decisions. The economic activity of a household determines key inputs (like income) which feed into the production function of health of a child. An example of this can be dependence of a household on nearby forest resources for fuel-wood consumption or for livelihood (if it sells these resources in a market). In this case the choice of use of fuel-wood by household affects the local pollution level in the region. Additionally the forest cover is affected by the demand for forest resources (like fuel-wood) in the market, which in turn affects the pollution level in the area where they are finally consumed. A similar logic holds true for crop residue burning as well, it is a conscious decision taken by a household which impacts local pollution levels and at the same time affects a farmer's income which is a determinant of child health. This essentially points towards the fact that local pollution level is endogenous in the region of economic activity of the household.

4.1 Identification

The household behavioural choice of collecting fuel-wood or crop-burning and household income are omitted variables in our specification hence the local pollution variable is en-

ogenous. To solve this endogeneity problem, we rely on an instrumental variable which in our case is the probability weighted number of upwind fire-events which take place in non-local areas. These fire-events refer to biomass burning events that include crop residue burning and forest fires. We use wind direction as well to tag each fire-event as an upwind or downwind event with respect to each cluster. An upwind fire-event refers to a fire-event from which wind is blowing towards the cluster and down-wind fire-events refer to events with wind blowing away from the cluster.

We use upwind fire-events which happen in non-local areas (between 75 and 100 km) as an IV as they impact local mean PM2.5 levels but are not affected by household behavioural choices. The IV that we use has been explained diagrammatically in Figure 3, where the light grey center denotes the cluster location, the white circle forms the 75 km radius around the cluster and the grey ring represents the area between 75 and 100 km radii around the cluster. Our endogenous variable is the mean PM2.5 variable which is calculated for the white circle (within 75 km) and the probability weighted number of upwind fire-events in the grey ring form the IV (between 75 and 100 km). We choose the fire-events in non-local far away areas to ensure that they belong to a region which is not a part of economic activity area of a household. This essentially removes the effect of dependence on crop-burning or nearby forest resources (or farmlands) for livelihood or fuel-consumption. By capturing fire-events in this ring, we ensure that we only capture the part which contributes to the local pollution levels but is not correlated with household behavioural choices. Exogenous changes in wind direction is unlikely to be correlated with local behavioural choices or economic activity in the area and thus non-local *upwind* fire-events serves as an ideal instrumental variable. Zheng et. al (2019) use a similar IV in their paper where they study the impact of air pollution on happiness levels using pollution level of neighbouring areas as an IV for local pollution levels.

One of the main limitations of our study is that we are only able to capture big fire-events as satellite recordings for fire-events fail to capture smaller fires which happen in smaller farms. The satellite can only capture fires which burn brightly enough (big enough to be quite intensive which guarantees that they get captured) to be recorded below the tree

canopy or cloud cover. The estimates we thus get can be considered as the lower-bound effect of outdoor pollution due to non-local fire-events on child health.

4.2 Fire-events Data

Our source of biomass burning events (called fire incidents) is NASA’s Fire Information for Resource Management System (FIRMS) data which captures real-time active fire locations across the globe. The FIRMS data that we use is called MODIS (shortform for MODerate Resolution Imaging Spectro radiometer) data and it records fire incidents at pixel level where each pixel is identified by a latitude and longitude reading. Each latitude (and longitude) is the center of a one kilometre fire pixel (1 km X 1 km in size). This data records not just the location of a fire but also the brightness (temperature) of fire (in Kelvin units) and date and time when the incident was picked by the Terra satellite. An observation for a fire incident in MODIS data for a latitude and longitude does not necessarily mean that the size of the fire is one square kilometre, but it means that atleast one fire is located within this fire pixel (under good conditions the satellite can detect fires as small as 100m^2). The MODIS data is available on a daily basis since November 2000 and NASA reports that the fires captured by this dataset are mostly vegetation fires. NASA data on fire incidents also provides a variable “confidence”, which depicts the quality of the observations and it ranges from 0-100. Following Vogl, 2016 we use this variable to construct a probability weighted count of fire-events around a cluster (between 75 and 100 km). NASA’s FIRMs data can also have some missing values attributable to satellite sensor outage however major incidents reported for sensor outage happened in years 2001-2003 which precedes our analysis period.

We use the cluster GIS information from DHS data and calculate the total number of all fire events which took place between 75 and 100 km radii (non-local exposure) during the first (in-utero) trimester of a child. To ensure respondent confidentiality, all clusters in the DHS data are displaced from their true location. The displacement is done by displacing an urban cluster by two kilometre and a rural cluster by five kilometre with one percent of the rural clusters being displaced by as much as 10 kilometres. The displacement can take place

in any direction but the cluster remains within the country boundary, within the same state and district. We take the radius for our analysis to be 75 kilometre which is large enough so that the true location of the cluster and sphere of economic activity of a household is contained within the 75 kilometre radius circle. Additionally capturing fire-events which take place between 75 to 100 kilometres radii ensures far away fire-incidents get captured (as smoke from fire-incidents which do not happen in the immediate neighbourhood can also travel long distances and affect local pollution levels).

Meteorological variables such as wind speed and wind direction are expected to play an important role in modulating the outflow of fire burning residues emitted from a fire event. To account for this, we tag each fire event with the respective wind speed and wind direction. We use ERA-Interim data of u (zonal wind) and v (meridional wind) at 10m from the European Centre for Medium-Range Weather Forecasts (ECMWF) ERA-Interim dataset at $0.125^*0.125$ degree resolution.

The wind speed and wind direction was estimated as in equation (2) and (3) respectively.

$$ws = \text{sqrt}[(u)^2 + (v)^2] \tag{2}$$

$$\text{winddirection} = [\text{atan}(u/v) * (180/\text{pi})] + 180 \tag{3}$$

Figure 4 provides a linear fit plot between local PM2.5 levels and non-local fire-events (all fire-events - left panel and just upwind fire-events - right panel). A strong positive relationship between the two is evident from this graph and forms the basis for using non-local fire-events as an IV for local pollution levels.

4.3 Fire-events in India

India has a substantial amount of land under cultivation(60%) and under forest cover(25%), with majority biomass burning events taking place in these areas. Over the past few decades, Indian agriculture has been marked with expansion of irrigation facilities, adoption of high yield variety seeds and increased mechanisation (like use of combine harvester). A combi-

nation of these factors led to adoption of multi-cropping system by farmers which leaves little time in between the harvest of one crop and sowing of another. In this scenario, crop residue burning thus emerged as the quickest and cheapest way to get the farm ready for the next crop. Cereals are the prime contributor to crop burning activity in India, with rice and wheat crop residue burning forming the major chunk of residue burning process (Jain et al, 2014). Two major residue burning seasons are thus related to crop harvest seasons: kharif crop harvest (rice stubble burning) which takes place in the months of October and November; and rabi crop harvest (wheat straw burning) which happens in the months of March to May.

Biomass burning in India is not limited to just crop residue burning, it covers forest fires as well. Forest fires or wildfires are caused by various factors acting in conjunction with each other. These factors include availability of biomass (dry vegetation) and appropriate climatic conditions (high temperature, low pressure, windy conditions). Forest Survey of India lists vulnerable months for each state when forest fires are most likely to happen, which mainly span the high temperature months from March to June. Wildfires happen due to both intentional and unintentional human activity. In North Eastern states and in states along the Eastern Ghats, slash and burn activity is rampant wherein vegetation in forests is cut (slashed) and then burned to clear the piece of land for human use. In a lot of cases unintentional human activities like leaving active cigarette butts behind in open forests lead to forest fires. Other natural factors which cause forest fires include lightening which produces a spark to start a fire in dry vegetation.

In Figure 5, we plot the temporal variation in PM2.5 and fire-events. The figure plots mean levels across all sampled clusters in the latest DHS round for India. We look at mean PM2.5, mean count of total fire-events which take place in non-local areas (75 to 100 km) and mean count of total upwind fire-events which take place in non-local areas (75 to 100 km). As shown in the graph (solid blue line) the winter months (from October to January) have highest pollution levels in comparison to summer months (March to June), with lowest pollution levels recorded in monsoon period (August-September). Corresponding to two harvest seasons we see two peaks in fire-events plots (both all fire-events and upwind fire-

events).

In western countries forest fires are mainly responsible for the carbon content release due to biomass burning; however, in case of India (and other South Asian countries) crop residue burning contributes the most to total carbon release. In South Asia, India stands out both in terms of total area burned (4.5 million hectares burned in 2015) and in terms of total carbon content (1.5 million metric tonnes) released due to biomass burning. A raw count of biomass burning events in India shows that roughly both crop residue burning and forest fires contribute equally. However, if we weigh these events based on the population density⁸ of the area in which these events occur then crop burning events contribute more to the total biomass burning events (65 percent). This mainly happens because residue burning activities happen in more populated areas as against forest fires which happen in low density areas. Appendix Figure 1 provides the population weighted split between forest fires and crop residue burning in few selected states in India. As can be seen in this graph, with an exception of Punjab, almost all other states are affected by both forest fires and residue burning.

Biomass burning is a major source of pollution as it releases harmful pollutants like Carbon Dioxide(CO₂), Carbon Monoxide (CO), Sulphur Oxides and particulate matter (PM) in the atmosphere. The release of harmful pollutants in the atmosphere is captured by aerosol loading⁹ in the region. Studies have found that aerosol loading increases in the downwind regions and in the vertical direction as well. Kaskaoutis et. al (2014) find that crop burning in Punjab has an effect on aerosol properties of the Indo-Gangetic Plains, also particulate matter (PM_{2.5}) concentrations increase near ground surface and the concentration of pollutants fall as we move from west to east India. To summarise, fine particulate matter released during biomass burning incidents have long range travel properties and affect not just the local areas but far away regions as well.

⁸ Geo-coded fire events have been projected onto land mask cover for India to categorise each fire event as an event which happens in a forest area vs cropped area. This data is then projected onto density map of India, to get the density of the population in which these events take place.

⁹ Aerosol loading is the suspensions of solids and/or liquid particles in the air that we breathe. Dust, smoke, haze are also part of aerosol loading.

5 Results

5.1 Pollution and Child Health

For our entire analysis, the main independent variable PM_{cdmt} and IV variable - probability weighted count of upwind fire-events in 75 to 100 kms have been converted into z-scores for the estimation sample. Table 2 presents the estimates for the first stage of 2SLS methodology outlined in the previous section. Columns 1 and 2 present the results for the first stage of our 2SLS methodology using two different Instrumental Variables, that is probability weighted total number of fire-events which lie between 75 to 100 km radius (IV1) and probability weighted total number of upwind fire-events which lie between 75 to 100 km radius (IV2). We introduce month, year and cluster fixed effects, month into year fixed effects and district into year fixed effects to account for any omitted variables at these levels. As expected the relationship between the endogenous variable - local PM2.5 in 75km radius and our IVs is positive and highly significant. A z-score increase in total fire-events leads to 0.16 standard deviation unit increase in PM2.5 (column 1). Similarly a one standard unit change in number of upwind fire-events leads to a 0.10 standard deviation unit increase in local pollution levels (column 2). This is in line with our hypothesis that particulate matter from fire-events far away affect local pollution levels. The first stage F-stat is also greater than 10 (rule of thumb) and rk-LM statistic is highly significant. These results represent that local PM2.5 variation is affected by the seasonality present in biomass burning events happening in non-local adjacent areas.

Table 3 presents the OLS and IV results of effect of mean outdoor pollution in the first trimester on child health outcomes. As explained in the previous section the OLS regression of outdoor pollution on child health (equation (1)) is riddled with endogeneity problem, hence the estimates that we see in columns 1 and 2 are biased. We next move to the second stage results obtained using 2SLS strategy. We find that both weight-for-age (WFA-Z) and height-for-age(HFA-Z) are negatively affected by outdoor pollution experienced in-utero during the first trimester. Columns 3 and 4 present our 2SLS results using all fire events as an IV. A standard deviation unit change in mean PM2.5 during first trimester leads to a decrease in

WFA-Z(HFA-Z) by $-0.068(-0.085)$ standard deviation units. Finally columns 5 and 6 provide 2SLS results using upwind fire-events in 75 to 100 kms as an IV. Using upwind fire-events increases the precision of our results and we find that a standard deviation unit change in mean PM_{2.5} during first trimester leads to a decrease in WFA-Z(HFA-Z) by $-0.102(-0.115)$ standard deviation units which translates into a 6.7 percent decrease in WFA-Z and 7.8 percent decrease in HFA-Z. These results are similar in magnitude to the results which were found by Brainerd and Menon(2014) owing to exposure to water pollutants in the month of conception.

Literature suggests that variation in concentration of pollutants is highly correlated as they often emanate from same sources. So although the focus of this study has been on outdoor pollution as captured by PM_{2.5} but the results that we see might represent the effect of other pollutants as well (like CO, CO₂ and Sulphur Oxides etc). And although we observe a negative relation between birth weight and exposure to pollution, the result is not significant unlike the Brazilian study by Vogl(2016) ¹⁰. In Appendix Table A2 we show the reduced form specification results (that is direct effect of non-local upwind fire events on child health indicators) for our estimation sample.

5.2 Robustness Checks

In this section we provide multiple robustness checks for our results. Local weather condition like rainfall can play an important role as rainfall makes the ash and other pollutant particles settle on the ground thereby reducing pollution levels. Temperature also plays an important role in pollution dynamics. We control for both local temperature and rainfall in Table 4 (columns 1 and 2). The number of observations is slightly smaller than before due to few random missing rainfall and temperature information. Our original results still hold and the magnitude of the effect is slightly larger after accounting for weather controls. The gestational age or pregnancy duration is also an important determinant of intrauterine growth of a foetus which affects future child health. We control for gestational age in column

¹⁰ Results available on request.

3 and 4 and find that our estimates remain unchanged.

The analysis upto now used upwind fire-events happening in 75 to 100 km radius as the IV for local mean PM2.5 in the 75 km radius around the cluster location. We now provide results for alternate radii specifications to test the sensitivity of our model. In Table 5, columns 1 and 2, the IV being used is the probability weighted total number of upwind fire-events in 50 to 100 km radius (compressing the white inner circle in Figure 5). In columns 3 and 4, the IV being used is the probability weighted total number of upwind fire-events in 50 to 75 km radius for local mean PM2.5 in the 50 km radius (compressing the donut in Figure 5). Reducing the local pollution radius to 50 kms leads to a significant drop in total number of observations as PM2.5 information is missing for a lot of observations. However we still find that our results are of similar magnitude (they are slightly smaller for HFA-Z analysis) and still remain significant. The HFA-Z result in Table 5 column 2 is significant at 10 percent level while in column 4 it is marginally significant at 10 percent level (p-value = 0.109). Lastly in columns 5 and 6, we drop the observations corresponding to the state of Punjab. This has been done to ensure that our results are not driven in any way by the state of Punjab which is affected by high levels of pollution corresponding to highest level of recorded fire-events in India¹¹. Our results become larger in magnitude and are more significant after dropping the state of Punjab.

We next provide some falsification tests for our analysis. In Table 6 (column 1 and 2) instead of using upwind fires as an IV we use downwind fires (which lie in opposite octant from that of upwind fires with wind blowing away from cluster location). We find that using downwind fire-events as an IV makes our results insignificant and in case of HFA-Z the sign also counter-intuitively turns positive. In columns 3 and 4, we provide results on the effect of pollution on child health where the location of a child has been randomly shuffled. This random assignment of location leads to counter-intuitive positive (and insignificant) effect of early life exposure to pollution during first trimester on child health which strengthens our hypothesis that location does matter when it comes to pollution exposure (and in turn affects child health).

¹¹ Almost 25% of total fire-events in India take place in Punjab.

We check whether our instrument meets exclusion restriction in Table 7. We regress various characteristics of a household (and its members) on our main IV - upwind fire-events, essentially an insignificant result shows that there is no systematic relationship between our IV and household (and its member's) characteristics. The level of education of mother and literacy status of father is not systematically related to our IV. The asset ownership of a household and rich-poor status of a household is also unrelated to our IV. The household size and vaccination behaviour is also not significantly related to fire-intensity in non-local areas.

Do mothers plan conception?

An important threat in our analysis can be avoidance behaviour by mothers, that is if mothers purposely avoid particular months for conception due to their concern about future child health related to seasonal biomass burning activities. We test this by looking at birth history of mothers for last five years, we calculate the number of conceptions in each month for each mother living in cluster c . For example, if a mother conceived two kids in two different years in the month of February then the count will be 2 for this variable corresponding to the month of February. We regress log of this variable (number of conceptions per month) on our IV i.e. non-local fire-events between 75 and 100 kms to see if there is any systematic relationship between the two. We also control for mother's education, father's literacy level, characteristics of household head and wealth index of the household. Month and cluster fixed effects are also introduced to account for any time related and region related heterogeneity. We present these results in Table 8. In column 1, we present results where we try and see whether there is correlation between three month exposure to upwind fire-events and log number of conceptions and in column 2, we assess whether exposure to fire-events in the month of conception has any effect on conception behaviour. We find that in both the cases non-local fire-events have no impact on mother's conception behaviour.

5.3 Heterogeneity

We provide disaggregated regressions for Height-for-age for the following sub-samples: Poor-Rich and North-South. By splitting our sample into Poor (wealth index lower than 2) and Rich sample (wealth index greater than equal to 3). In Table 9, we find that the effect is present only for poor households. This can possibly be due to the fact that children in poor households have less access to health care to abate negative effect of pollution on health. We also provide results by splitting our sample into Northern-Southern states and find that most of the effects that we see are limited to North India.

We now focus on other time windows of critical development, that is second, third trimester and the post-natal period of first three months after birth. Table 10 summarises our results, we find that in-utero exposure to outdoor pollution which is experienced by the mother (and her foetus) for second, third trimester and post-natal period has no impact on Height-for-age, but some negative effect is present for Weight-for-age corresponding to exposure in second trimester.

6 Conclusion

Outdoor pollution in India breaches safe standards in many areas. We link outdoor pollution to biomass burning which is a significant source of carbonaceous aerosols, it plays a vital role in atmospheric chemistry, air quality, ecosystems, and human health. Our analysis shows that outdoor pollution is affected by neighbouring biomass burning events; this is used to causally infer the effect of outdoor pollution (as measured by PM_{2.5}) on child growth indicators. We find that a z-score increase in PM_{2.5} levels during first trimester leads to a reduction in Height-for-age (HFA-Z, stunting measure) and Weight-for-age (WFA-Z, underweight measure) by 0.115 and 0.102 standard deviation units respectively. Figure 6 summarises our results graphically, exposure to outdoor pollution during different critical windows of growth of a child is associated with worse child health outcomes. Almost all the estimates are negative with significant effect present for exposure to pollution during first trimester and second trimester (only WFA-Z measure).

The above results establish that exposure to pollution is linked to stunting measure (HFA) in childhood. We now provide an estimate of this problem on GDP of India using a back-of-an-envelope calculation based on the Galasso et al. (2016) study. This study does a literature review of the effect of stunting on GDP. Stunting affects GDP of a nation via three channels: lower returns to lower education, lower returns to lower height and lower returns to lower cognition. For India, where 66 percent of the workforce was stunted in childhood, this study estimates that a complete elimination of stunting would have increased GDP by 10 percent ¹². We use a point estimate of probability of being stunted due to outdoor pollution, and find that one standard deviation increase in outdoor pollution leads to a 0.18 percent reduction in GDP.

India needs effective policies regarding regulation and management of outdoor pollution, since the current policies are ineffective. Cross-border policies are needed to tackle the problem of pollution. To curb air pollution, effective management of forest fires is needed; however, the budget allocation for this purpose is really small and remains unused in every financial year. Similarly the government has committed itself to subsidising the use of happy-seeder technology (this is an alternative to combine harvester, it leaves rice residue in form of a mulch on farm which doesn't hamper wheat crop sowing and hence doesn't require burning), however the uptake of this policy remains quite low due to high initial investment in the machine (Gupta and Somnathan, 2016). The National Clean Air Program (2018) is a welcome step in this domain as it plans to extend air quality monitoring network, conduct intensive awareness and monitoring campaigns, create city-specific action plans, among many other initiatives.

¹² This is an average figure for South Asia.

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Tables and Figures

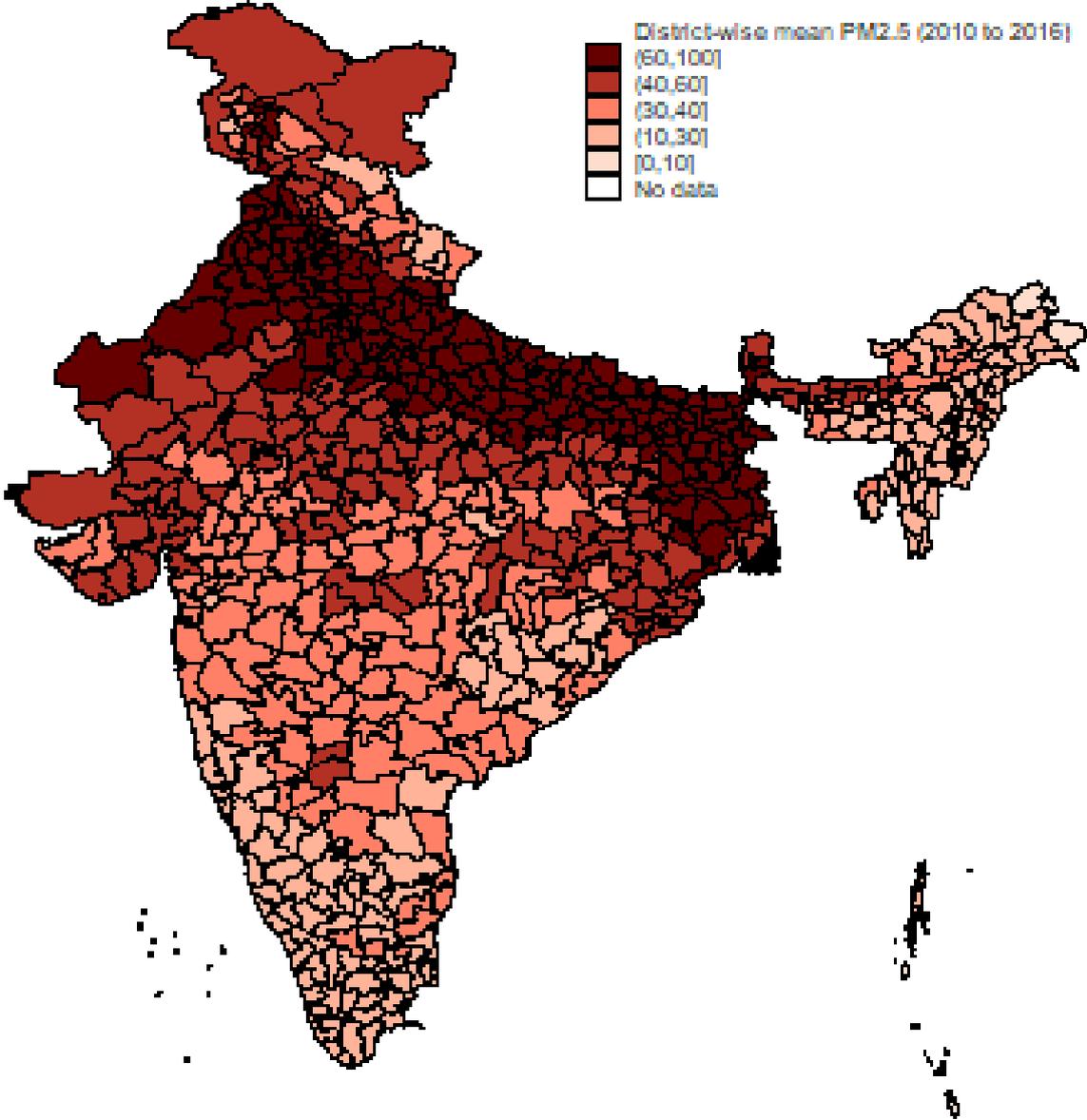


Figure 1: Spatial variation in Pollution: Mean PM2.5 in districts of India (2010 to 2016)

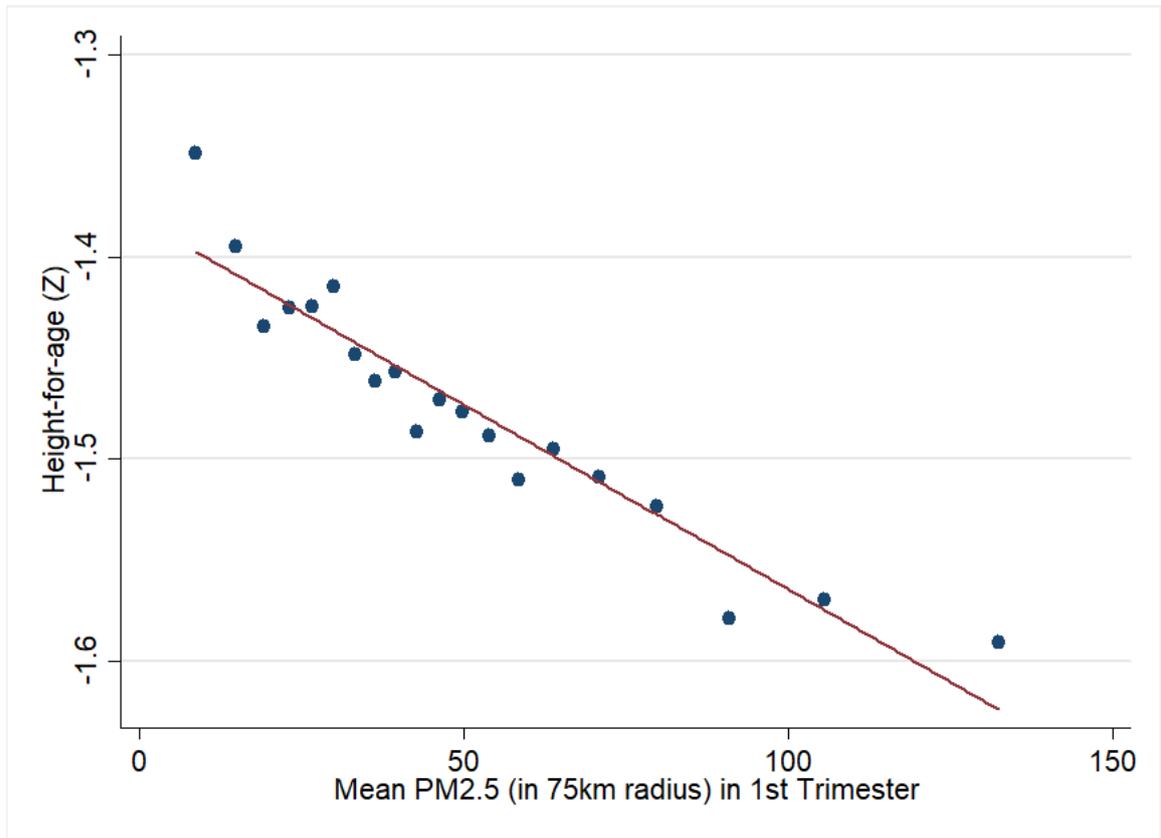


Figure 2: Binscatter plot for relationship between Height-for-age Z scores and Mean PM2.5 in first trimester.

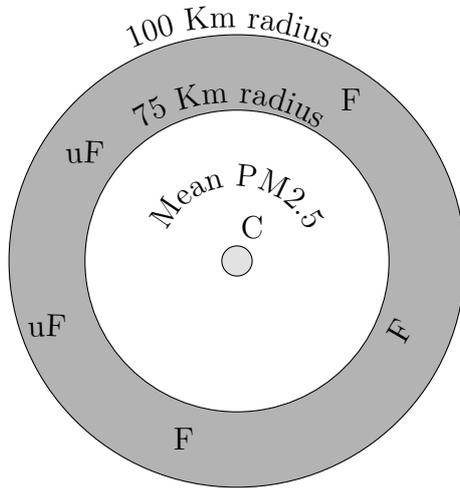


Figure 3: Identification Strategy: Center (smallest grey circle) represents the cluster location, White circle corresponds to 75km radius circle around the cluster location, grey ring area (donut shape) corresponds to area between two circles (75 and 100 Km radii circles) with cluster location as the center. Mean pollution level is calculated for the white circle, we call this *local* pollution level for cluster C.

Local pollution level is instrumented using i) *non-local* biomass burning events which take place in the grey ring area (F + uF). ii) Upwind *non-local* biomass burning events which take place in the grey ring area (only uF). Probability weighted counts used everywhere.

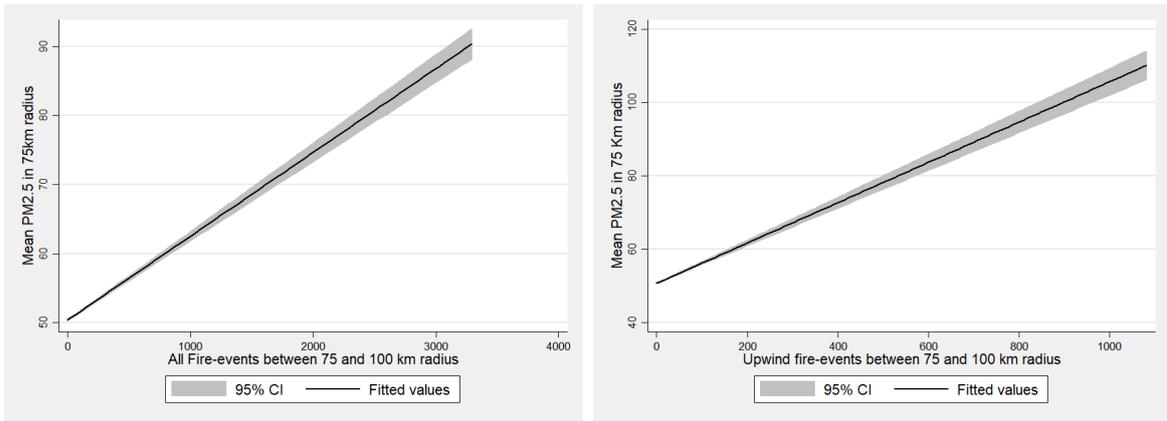


Figure 4: Linear fit plot between Mean PM2.5 (in 75 km radius), Total number of fire-events between 75 and 100 km radius(Non-local fire-events) & Total number of upwind fire-events between 75 and 100 km radius (Non-local upwind fire-events). Unit of observation is a child, shaded area is 95% confidence interval.

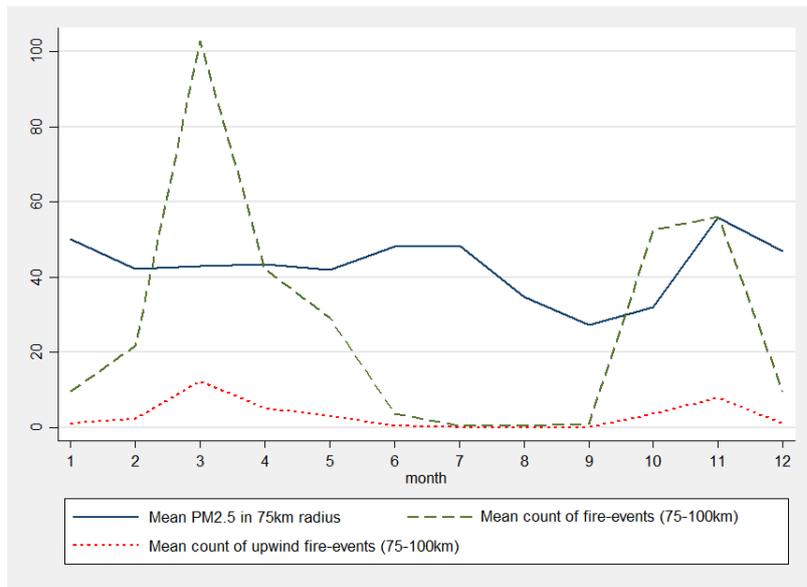


Figure 5: Mean PM2.5, Mean count of all fire-events (in 75-100 km radius) & Mean count of upwind fire-events (in 75-100 km radius) for each month in every year from 2010 to 2016. Figure represents mean over all sampled clusters belonging to all states of India.

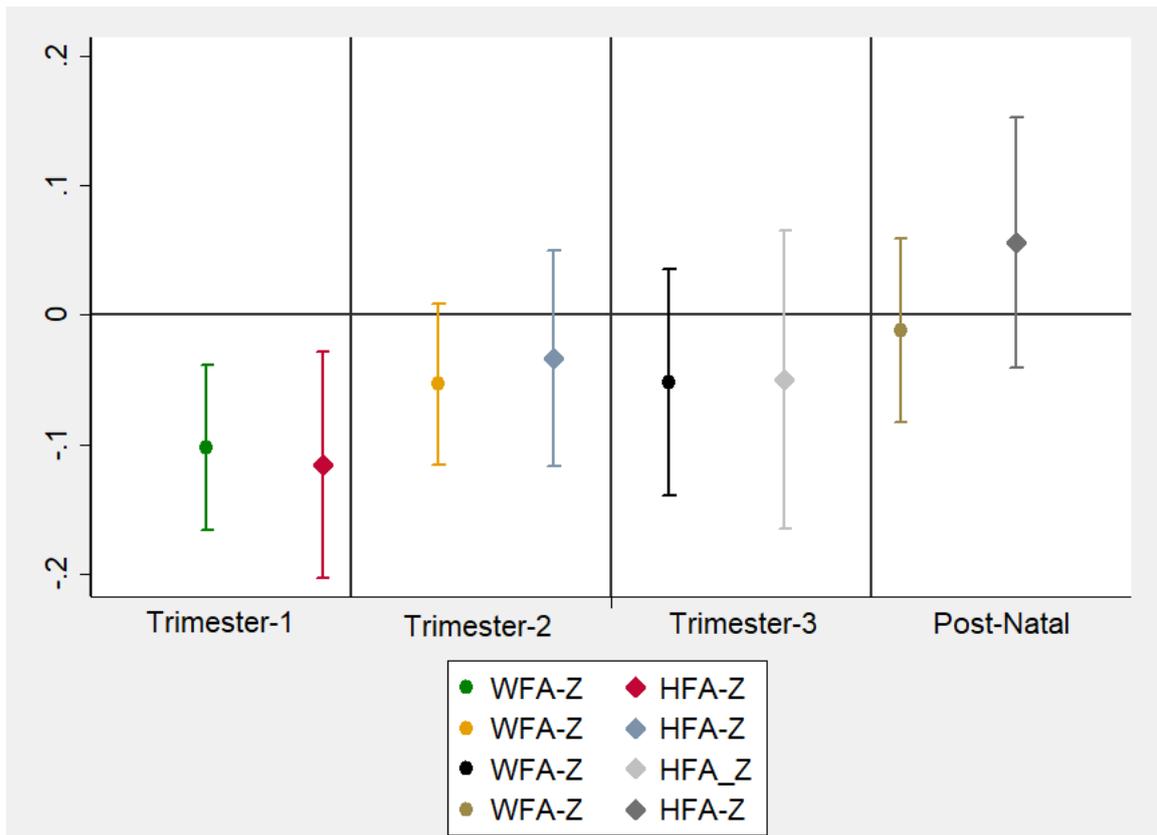


Figure 6: Coefficient of 2SLS regression of outcomes(HFA-Z and WFA-Z) on outdoor air pollution for different critical windows of development of a child. Vertical lines represent 95 percent confidence intervals.

Table 1: Summary Statistics

Variable	All India		NorthIndia		SouthIndia	
	Mean	Std Dev	Mean	Std Dev	Mean	Std Dev
<i>Outcomes</i>						
Height-for-age z score	-1.46	1.68	-1.52	1.68	-1.27	1.66
Weight-for-age z score	-1.52	1.22	-1.54	1.22	-1.44	1.22
Weight-for-height z score	-0.97	1.38	-0.96	1.38	-1.01	1.41
<i>Child characteristics</i>						
Dummy for male child	0.52	0.50	0.52	0.50	0.51	0.50
Birth-order	2.26	1.46	2.38	1.54	1.88	1.05
Childage in months	29.00	16.55	29.02	16.58	28.94	16.46
Pregnancy duration	9.02	0.48	9.00	0.46	9.10	0.52
<i>Parents characteristics</i>						
Mother's age at birth	24.50	4.91	24.67	5.00	23.91	4.56
Mother's number of education years	6.26	5.16	5.81	5.18	7.81	4.77
Dummy for literate father	0.97	0.17	0.97	0.18	0.98	0.15
<i>Household characteristics</i>						
Rural	0.76	0.43	0.77	0.42	0.70	0.46
Dummy for head of household being a male	0.88	0.33	0.88	0.33	0.89	0.32
Age of head of household	44.59	15.18	44.51	15.21	44.88	15.09
Dummy for source of water: Pipedwater	0.23	0.42	0.23	0.42	0.24	0.43
Dummy for using clean cooking fuel	0.28	0.45	0.26	0.44	0.37	0.48
Dummy for household practicing open defecation (OD)	0.42	0.49	0.41	0.49	0.44	0.50
Fraction of HHs practicing OD in a village	0.43	0.35	0.43	0.35	0.45	0.33
Has electricity connection	0.85	0.36	0.82	0.38	0.94	0.23
Religion = Hindu	0.73	0.44	0.70	0.46	0.83	0.38
Household size	6.57	2.87	6.75	2.93	5.99	2.57
<i>Mean pollution in 75 km radius</i>						
1st Trimester	54.06	31.87	59.33	32.90	35.96	19.07
2nd Trimester	52.23	29.84	56.98	30.90	35.90	18.02
3rd Trimester	53.30	32.40	58.52	33.68	35.32	18.47
Post-natal (3 months after birth)	53.32	35.38	58.46	36.89	35.44	21.46
Observations	1,81,361		1,40,476		40,885	

Table 2: First-stage regression

	Mean PM2.5 in 75 km radius(Z)	
	(1)	(2)
IV1: Number of fire events between 75 and 100 km radius (Z)	0.165*** (0.003)	
IV2: Number of upwind fire events between 75 and 100 km radius (Z)		0.105*** (0.003)
First Stage F-stat	2099	863
rk LM statistic	1997	1005
Anderson Rubin wald statistic (p-value)	0.0018	0.0016
Observations	179816	179816
Includes other controls from 2nd stage	Yes	Yes
Includes FEs for Month, Year & Cluster	Yes	Yes
Includes FEs for Month*Year & District*Year	Yes	Yes

Note: Each coefficient corresponds to an individual FIRST stage 2SLS regression of HFA-Z or WFA-Z on variables mentioned in the first column. Standard errors in parentheses are clustered by DHS cluster. Notation for p-values *** is $p < 0.01$, ** is $p < 0.05$ & * is $p < 0.1$. Regressions include other controls - gender, birth order and age of child, mother's years of education, mother's age at birth and its square, age and gender of household head, dummy for whether household has pipedwater, has clean cooking source, whether household practices open defecation and fraction of the village who practice open defecation(excluding self). PM2.5, Fire-events variables (IV1, IV2) have all been converted into z-scores.

Table 3: Instrumental variable regression of outcomes on weighted PM2.5 (z-score)

	OLS		IV1		IV2	
	(1)	(2)	(3)	(4)	(5)	(6)
	WFA-Z	HFA-Z	WFA-Z	HFA-Z	WFA-Z	HFA-Z
Trimester-1: Mean PM2.5 in 75km radius (Z)	-0.0121**	-0.0116	-0.0683***	-0.0854***	-0.102***	-0.115***
	(0.00512)	(0.00709)	(0.0219)	(0.0298)	(0.0323)	(0.0447)
Mean of Dependent Variable	-1.52	-1.46	-1.52	-1.46	-1.52	-1.46
Includes Child, Mother & HH characteristics	Yes	Yes	Yes	Yes	Yes	Yes
Includes FEs for Month, Year & Cluster	Yes	Yes	Yes	Yes	Yes	Yes
Includes FEs for Month*Year & District*Year	Yes	Yes	Yes	Yes	Yes	Yes
Observations	179816	179816	179816	179816	179816	179816

Note: Standard errors in parentheses are clustered by DHS cluster. Notation for p-values *** is $p < 0.01$, ** is $p < 0.05$ & * is $p < 0.1$. Each coefficient corresponds to an individual OLS or 2SLS regression of HFA-Z or WFA-Z on weighted mean PM2.5 in first trimester (z-score). Regressions include controls which are same as those mentioned in Table 2 notes.

Table 4: Robustness Checks: Extended control for weather & gestational age

	IV: Upwind fire-events b/w 75 and 100 km radius			
	(1)	(2)	(3)	(4)
	WFA-Z	HFA-Z	WFA-Z	HFA-Z
Trimester-1: Mean PM2.5 in 75km radius (Z)	-0.111*** (0.0328)	-0.125*** (0.0453)	-0.103*** (0.0323)	-0.116*** (0.0448)
Mean rainfall in 75 km radius	-0.046** (0.021)	-0.049 (0.030)		
Mean temperature in 75 km radius	0.003** (0.001)	0.003 (0.002)		
Gestational period			0.137*** (0.025)	0.165*** (0.034)
Observations	179459	179459	179816	179816
Includes Child, Mother & HH characteristics	Yes	Yes	Yes	Yes
Includes FEs for Month, Year & Cluster	Yes	Yes	Yes	Yes
Includes FEs for Month*Year & District*Year	Yes	Yes	Yes	Yes

Note: Standard errors in parentheses are clustered by DHS cluster. Notation for p-values *** is $p < 0.01$, ** is $p < 0.05$ & * is $p < 0.1$. Each coefficient corresponds to an individual 2SLS regression of HFA-Z or WFA-Z on weighted mean PM2.5 in first trimester (z-score). Regressions include controls which are same as those mentioned in Table 2 notes.

Table 5: Sensitivity Analysis

	IV: Upwind Fire events b/w 50 and 100 Km radius	(1)	(2)	(3)	(4)	(5)	(6)
	WFA-Z	HFA-Z	WFA-Z	HFA-Z	WFA-Z	HFA-Z	HFA-Z
Trimester-1: Mean PM2.5 in 50km radius (Z)	-0.101*** (0.0331)	-0.0867* (0.0461)	-0.111*** (0.0349)	-0.0770 (0.0481)			
Trimester-1: Mean PM2.5 in 75km radius (Z)							
Observations	164462	164462	164462	164462	175141	175141	175141
Includes Child, Mother & HH characteristics	Yes	Yes	Yes	Yes	Yes	Yes	Yes
Includes FEs for Month, Year & Cluster	Yes	Yes	Yes	Yes	Yes	Yes	Yes
Includes FEs for Month*Year & District*Year	Yes	Yes	Yes	Yes	Yes	Yes	Yes

Note: Standard errors in parentheses are clustered by DHS cluster. Notation for p-values *** is $p < 0.01$, ** is $p < 0.05$ & * is $p < 0.1$. Each coefficient corresponds to an individual 2SLS regression of HFA-Z or WFA-Z on weighted mean PM2.5 in first trimester (z-score). Regressions include controls which are same as those mentioned in Table 2 notes.

Table 6: Falsification Tests

	IV: Downwind Fire-events in 75 to 100km radius		IV: Upwind Fire-events in 75 to 100 km radius (shuffled location)	
	(1)	(2)	(3)	(4)
	WFA-Z	HFA-Z	WFA-Z	HFA-Z
Trimester-1: Mean PM2.5 in 75km radius (Z)	-0.008 (0.03)	0.003 (0.04)	0.033 (0.033)	0.054 (0.048)
Observations	179816	179816	179539	179539
Includes Child, Mother & HH characteristics	Yes	Yes	Yes	Yes
Includes FEs for Month, Year & Cluster	Yes	Yes	Yes	Yes
Includes FEs for Month*Year & District*Year	Yes	Yes	Yes	Yes

Note: Standard errors in parentheses are clustered by DHS cluster. Notation for p-values *** is $p < 0.01$, ** is $p < 0.05$ & * is $p < 0.1$. Each coefficient corresponds to an individual 2SLS regression of HFA-Z or WFA-Z on weighted mean PM2.5 in first trimester (z-score). Regressions include controls which are same as those mentioned in Table 2 notes.

Table 7: IV Validity

	Mother's Education	Father's Literate	Asset Ownership	Poor	Vaccination	HH size
	(1)	(2)	(3)	(4)	(5)	(6)
Upwind fire events between 75 and 100km radius in 1st Trimester (Z)	0.016 (0.01)	-0.0002 (0.0003)	-0.001 (0.002)	0.001 (0.0008)	0.0007 (0.001)	0.01 (0.006)
Observations	179816	179816	179816	179816	179816	179816
Includes Child, Mother & HH characteristics	Yes	Yes	Yes	Yes	Yes	Yes
Includes FEs for Month, Year & Cluster	Yes	Yes	Yes	Yes	Yes	Yes
Includes FEs for Month*Year & District*Year	Yes	Yes	Yes	Yes	Yes	Yes

Standard errors in parentheses are clustered by DHS cluster. Notation for p-values *** is $p < 0.01$, ** is $p < 0.05$ & * is $p < 0.1$. Fire-events have been converted to z-scores.

Table 8: Mother's conception behaviour

	(1)	(2)
	log_conceptions	log_conceptions
3 month exposure to Upwind fire-events in 75-100kms	0.00732 (0.0119)	
Exposure to Upwind fire-events in 75-100kms in the month of conception		0.0119 (0.0219)
Observations	197753	197753
Cluster FE	Yes	Yes
Month FE	Yes	Yes

Note: Standard errors in parentheses are clustered by DHS cluster. Notation for p-values *** is $p < 0.01$, ** is $p < 0.05$ & * is $p < 0.1$. Each coefficient corresponds to an individual OLS regression of log number of conceptions on controls mentioned in the table. Regressions include other controls for mother's education, father's literacy level, characteristics of household head and wealth index of the household.

Table 9: Heterogeneity

	IV regression: Height-for-age Z score			
	Poor (1)	Rich (2)	North (3)	South (4)
Trimester-1: Mean PM2.5 in 75km radius(Z)	-0.139** (0.0592)	-0.0824 (0.0633)	-0.120*** (0.0433)	-0.0226 (1.813)
Observations	84458	88932	139656	40159
Includes Child, Mother & HH characteristics	Yes	Yes	Yes	Yes
Includes FEs for Month, Year & Cluster	Yes	Yes	Yes	Yes
Includes FEs for Month*Year & District*Year	Yes	Yes	Yes	Yes

Note: Standard errors in parentheses are clustered by DHS cluster. Notation for p-values *** is $p < 0.01$, ** is $p < 0.05$ & * is $p < 0.1$. Each coefficient corresponds to an individual 2SLS regression of HFA-Z on weighted mean PM2.5 in first trimester (z-score). Regressions include controls which are same as those mentioned in Table 2 notes.

North Indian states: Arunachal Pradesh, Assam, Bihar, Chandigarh, Gujarat, Haryana, Himachal Pradesh, Jammu and Kashmir, Jharkhand, Madhya Pradesh, Manipur, Meghalaya, Mizoram Nagaland, Delhi, Punjab, Rajasthan, Sikkim, Tripura, Uttar Pradesh and Uttarakhand.

South Indian states: Andhra Pradesh, Karnataka, Kerala, Maharashtra, Chhattisgarh, Odisha, Telangana, West Bengal, Lakshwadeep Islands, Andaman and Nicobar Islands, Dadar and Nagar Haveli, Daman and Diu, Puducherry and Goa.

Table 10: Instrumental variable effects:

Impact of weighted PM2.5 in 2nd Trimester to Post-natal period

	(1)	(2)
	WFA-Z	HFA-Z
Trimester-2: Mean PM2.5 in 75km radius (Z)	-0.05*	-0.03
	(0.03)	(0.04)
Observations	184183	184183
Trimester-3: Mean PM2.5 in 75km radius (Z)	-0.05	-0.05
	(0.04)	(0.05)
Observations	172917	172917
Post-natal: Mean PM2.5 in 75km radius (Z)	-0.01	0.05
	(0.03)	(0.04)
Observations	190717	190717
Includes Child, Mother & HH characteristics	Yes	Yes
Includes FEs for Month, Year & Cluster	Yes	Yes
Includes FEs for Month*Year & District*Year	Yes	Yes

Note: Standard errors in parentheses are clustered by DHS cluster. Notation for p-values *** is $p < 0.01$, ** is $p < 0.05$ & * is $p < 0.1$. Each coefficient corresponds to an individual 2SLS regression of HFA-Z or WFA-Z on weighted mean PM2.5 in first trimester (z-score). Regressions include controls which are same as those mentioned in Table 2 notes.

Appendix

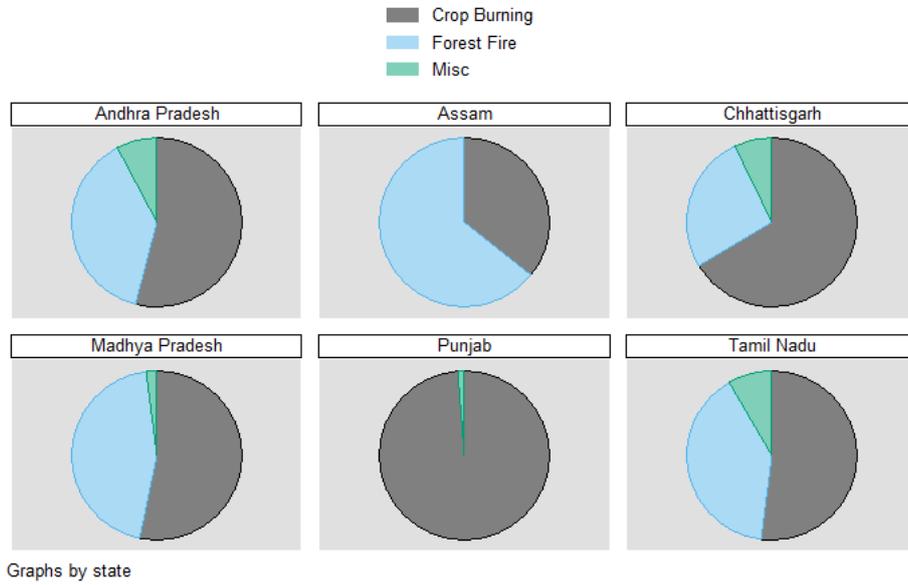


Figure 1: Population weighted split of all biomass burning events which took place from 2010-2016 for select states.

Table A1: Missing PM2.5 analysis

	(1)	(2)	(3)	(4)
	HFA-Z	Mother's	HH size	Father literate
	Education			
Dummy for Missing PM2.5	-0.003 (0.013)	-0.004 (0.033)	-0.013 (0.019)	-0.0006 (0.001)
Observations	223150	223150	223150	223150
Includes Child, Mother & HH characteristics	Yes	Yes	Yes	Yes
Includes FEs for Month, Year & Cluster	Yes	Yes	Yes	Yes
Includes FEs for Month*Year & District*Year	Yes	Yes	Yes	Yes

Note: Standard errors in parentheses are clustered by DHS cluster.

Notation for p-values *** is $p < 0.01$, ** is $p < 0.05$ & * is $p < 0.1$.

Table A2: Reduced form Effect

	(1)	(2)
	WFA-Z	HFA-Z
Trimester-1: Upwind Fire events (75 to 100 kms radius)	-0.00845** (0.00414)	-0.0156*** (0.00560)
Trimester-2: Upwind Fire events (75 to 100 kms radius)	0.00439 (0.00418)	-0.00453 (0.00550)
Trimester-3: Upwind Fire events (75 to 100 kms radius)	0.00303 (0.00410)	-0.00568 (0.00526)
Includes Chils, Mother & HH Characteristics	Yes	Yes
Includes FE for Month, Year, Cluster	Yes	Yes
Includes FE for Month*Year & District*Year	Yes	Yes
Observations	179449	179449

Note: Standard errors in parentheses are clustered by DHS cluster. Notation for p-values *** is $p < 0.01$, ** is $p < 0.05$ & * is $p < 0.1$. Each coefficient corresponds to an individual OLS regression of HFA-Z or WFA-Z on fire-events (z-score). Regressions include controls which are same as those mentioned in Table 2 notes.

QUALITY

INDEPENDENCE

IMPACT

BROOKINGS INDIA